

Available online at www.sciencedirect.com



Journal of Photochemistry Photobiology A:Chemistry

Journal of Photochemistry and Photobiology A: Chemistry 179 (2006) 224-228

www.elsevier.com/locate/jphotochem

Mechanism of phenol decomposition on Fe–C–TiO₂ and Fe–TiO₂ photocatalysts via photo-Fenton process

Beata Tryba^{a,b,*}, Antoni W. Morawski^a, Michio Inagaki^c, Masahiro Toyoda^b

^a Technical University of Szczecin, Chemical Technology and Environmental Engineering, Ul. Pulaskiego 10, 70-322 Szczecin, Poland ^b Oita University, 700 Dannoharu, Oita 870-1192, Japan

^c Aichi Institute of Technology, Yakusa, Toyota 470-0392, Japan

Received 18 May 2005; received in revised form 22 July 2005; accepted 12 August 2005 Available online 23 September 2005

Abstract

Photoactivity of TiO_2 photocatalyst for phenol decomposition was enhanced through its modification by the mounting of iron. Two types of $Fe-C-TiO_2$ photocatalysts were prepared, the one having the high content of carbon, around 7.5–15.5 mass% (obtained from carbonization of poly(ethylene terephthalate) (PET)), and the other, contained the residual carbon, 0.2–3.3 mass% (obtained from the carbonization of oxalate). Photodecomposition of phenol on these $Fe-C-TiO_2$ photocatalysts proceeded through the photo-Fenton reactions and was faster on the photocatalysts with the residual carbon, which had high content of iron. The adsorption of phenol was higher on the photocatalysts having the high content of carbon. On $Fe-TiO_2$, the phenol decomposition was accelerated insignificantly by comparison with the pristine TiO_2 . It was proved that different mechanisms of phenol decomposition occurred on $Fe-TiO_2$ and $Fe-C-TiO_2$, on the former the phenol was decomposed by the radical reaction and on the latter by the complex reaction with the iron and intermediates of phenol decomposition. The radical reaction on $Fe-TiO_2$ photocatalysts, the rate of phenol decomposition is sped up significantly, resulting in the complex reactions of phenol in a very short time. © 2005 Elsevier B.V. All rights reserved.

Keywords: Phenol decomposition; Photo-Fenton reaction; Carbon-coated TiO₂; FeC₂O₄

1. Introduction

Carbon-coated TiO₂ photocatalysts were recently developed and studied [1–6]. It was proved that carbon-coated TiO₂ had higher photoactivity for methylene blue decomposition in water than pristine TiO₂ [1–3,6]. Some advantages of carbon coating were found such as suppressing of phase transformation from anatase to rutile during heat-treatment at high temperatures, increasing the adsorption of the organic compounds on the photocatalyst surface and prevention from the interaction between TiO₂ particles and the organic binder during photocatalytic decomposition [1–6]. However, the phenol decomposition on carbon-coated TiO₂ photocatalyst could not be enhanced, therefore a chemical modification of the carbon layer was performed. It was already reported that carbon-coated TiO₂ photocatalyst

1010-6030/\$ - see front matter © 2005 Elsevier B.V. All rights reserved. doi:10.1016/j.jphotochem.2005.08.019 modified by iron could successfully decompose phenol via photo-Fenton process [7].

Photo-Fenton reactions are widely studied and used in the photocatalytic process [8–15]. The most commonly used catalyst in the photo-Fenton process is iron [8]. In the presence of Fe^{2+} and H_2O_2 phenol decomposition is going very fast with and without light. When iron exists only in the form of Fe^{3+} , Fenton reaction is limiting by the reduction of Fe^{3+} to Fe^{2+} in the presence of H_2O_2 , which is less efficient than oxidation of Fe^{2+} , but after build-up of enough amount of Fe^{2+} the significant acceleration of degradation is obtained [12].

It was reported that the addition of hydroquinone or catehol to the reaction solution could reduce Fe^{3+} to Fe^{2+} , accelerating in this way decomposition of phenol by the formation of ternary HQ–Fe–H₂O₂ [12].

It was also reported that in the degradation of phenol via photo-Fenton process, iron complex plays an important role, phenol is more likely degraded through the complex oxidation than oxidation by hydroxyl radical [13].

^{*} Corresponding author. Tel.: +48 91 4494277; fax: +48 91 4494686. *E-mail address:* beata.tryba@ps.pl (B. Tryba).

In the present study, the comparison of phenol decomposition on Fe–C–TiO₂ and Fe–TiO₂ photocatalysts is reported. Some advantages of carbon coating on TiO₂ were found for mounting iron and its application as the photocatalyst in the photo-Fenton process.

2. Experimental

Four types of TiO₂ photocatalysts were prepared. Their preparation conditions are summarized in Table 1.

TiO₂ as-received, commercially available anatase powder ST-01 (Ishihara Co. Ltd., Japan), was used as the reference photocatalyst and also used as the pristine material for other types of modified TiO₂. It was known to consist of pure anatase phase with the crystallite size of about 7 nm and the aggregates of the size of about 60 nm and to have relatively high surface area as $300 \,\mathrm{m^2/g}$.

Fe-modified TiO₂, Fe-TiO₂, was prepared by heating the powder mixture of ST-01 with FeC₂O₄ in air at 550 $^{\circ}$ C for 3 h, the selected mixing ratio was 10/1 in mass.

Two types of Fe-C-TiO₂ photocatalysts were prepared. Fe-C-TiO₂ with the residual carbon were prepared from anatase-type TiO₂ (ST-01, Ishihara Sangyo Co. Ltd, Japan) and FeC₂O₄. Suspended TiO₂ in FeC₂O₄ solution (TiO₂/FeC₂O₄ ratio = 10/1 g/g) was mixed with magnetic stirrer firstly at room temperature for around 20 h, and then until the water completely evaporated at 100 °C. After drying in oven through overnight, modified TiO₂ thus prepared was heated at temperatures from 400 to 900 °C in Ar for 1 h. For preparation of Fe–C–TiO₂ photocatalysts with high carbon content, PET was added to the water solution of FeC₂O₄ in the mass ratio of PET/FeC₂O₄ = 10:1 and that mixture was stirred at room temperature for 20 h and then was heated up to 100 °C to evaporate the water. After drying in oven modified PET was mechanically mixed with TiO2 in the mass ratio 3:7 respectively and was heat-treated at 400-800 °C in Ar for 1 h [7]. Fe-TiO₂ photocatalyst was prepared by mixing the powders of TiO_2 with FeC_2O_4 in the mass ratio of $TiO_2/FeC_2O_4 = 10:1$ and heating in air at 550 °C for 3 h.

The prepared photocatalysts were characterized by BET surface area, XRD, FT-IR and the contents of iron and carbon. BET surface area was determined from the adsorption isotherm of nitrogen at 77 K. XRD analysis was performed with using Cu K α radiation. Content of iron in the catalysts (Fe²⁺ and Fe³⁺) was determined by extraction with 6 M HCl of 5 ml for 2 h [16] and then analyzed in the UV-vis spectrophotometer by using colorimetric method. Iron detection was carried out using 2,4,6tris(2-pyridyl)-1,3,5-triazine reagent with very high sensitivity to Fe²⁺. Content of carbon in Fe–C–TiO₂ photocatalysts was

Table 1 The preparation conditions of samples

The preparation conditions of samples						
Sample	TiO ₂ /FeC ₂ O ₄	TiO ₂ /PET	Heat treatment			
TiO ₂ (ST-01)	_	_	_			
Fe-TiO ₂	10/1 in powder	_	550 °C in air fo			

2 ()			
Fe-TiO ₂	10/1 in powder	_	550 °C in air for 3 h
Fe-C-TiO2	10/1 in water	_	400-900 °C in Ar for 1 h
Fe-PET-TiO2	10/1 in water	3/7	400-800 °C in Ar for 1 h

measured by TG analysis, combustion of carbon in air by heating the sample in air atmosphere up to $1000 \,^{\circ}$ C.

Photoactivity of the catalysts was measured for the phenol decomposition in water. Photocatalysts in the mass of 0.2 g/L were added to the 500 mL of phenol solution with concentration of 2.1×10^{-4} mol L⁻¹ in the "batch mode" reactor and that mixture was stirred for 3 h in the dark until saturation of adsorption. After adsorption, photodecomposition of phenol was performed under UV irradiation. Reactor was equipped with three black light blue fluorescent lamps of 20 W power. These lamps irradiate the light at wavelength range of 300-415 nm with maximum intensity at 352 nm. Irradiation power of UV rays inside the reactor was $2.4 \,\mathrm{mW}\,\mathrm{cm}^{-2}$ in the wavelength range of $310-400 \,\mathrm{nm}$ and 0.19 mW cm^{-2} in the wavelength range of 360–470 nm. The value of pH of the reaction mixtures was around 6.1. Some experiments at pH 2.8 were also performed. Experiments were carried out under UV with or without the addition of $0.03 \text{ mol/L H}_2O_2$ to the phenol solution. The changes in the phenol concentration and formation of hydroquinone (HQ), benzoquinone (BQ), catehol and short-chain compounds as the decomposition products were measured by high performance liquid chromatography (HPLC).

3. Results

The characteristics of TiO₂ and Fe–C–TiO₂ photocatalysts prepared with high amount of carbon are presented in Table 2.

Those samples consisted from a single anatase phase and had 7.5-15.5 mass% of carbon, higher for the lower temperature of heat-treatment. BET surface area did not differ much in those samples. The high capacity for phenol adsorption on Fe-PET400-TiO₂ was caused by the difference in the chemical nature of the carbon surface as reported in our previous paper [7].

The phenol decomposition performed at pH 6.1 under UV irradiation with the addition of H₂O₂ on TiO₂ and Fe-C-TiO₂ photocatalysts with high carbon content is presented in Fig. 1.

Only sample heat-treated at 400 °C, which showed high adsorption of phenol, showed high rate of phenol decomposition under UV with H_2O_2 , much higher than the pristine TiO₂ [7].



Fig. 1. Decomposition of phenol under UV irradiation and with H₂O₂ on TiO₂ and Fe-C-TiO₂ photocatalysts with high carbon content.

Table 2

HTT (°C)	Sample code	Content of TiO ₂ (mass%)	BET surface area (m ² /g)	Phases (by XRD)	Amount of Fe (mass%)		Phenol adsorption (10^{-5} mol/g)
					Fe ²⁺	Fe ³⁺	-
0	TiO ₂	100	300	А	0	0	1.0
400	Fe-PET400-TiO2	84.5	143	А	0.33	0.01	29.2
500	Fe-PET500-TiO2	89.5	150	А	0.44	0.02	2.7
600	Fe-PET600-TiO2	90	143	А	0.43	0.07	4.2
700	Fe-PET700-TiO2	91	142	А	0.6	0.07	3.7
800	Fe-PET800-TiO2	92.5	128	А	0.46	0	5.3

Characteristics of pristine TiO₂ and Fe-C-TiO₂ with high carbon content

Phenol decomposition performed at pH 2.8 insignificantly accelerated the rate of phenol decomposition (the results are not presented here), however the leaching of iron from the photocatalysts was noticed, what was not convenient, so all the other experiments were performed at neutral pH.

In order to investigate the role of the phenol adsorption on the surface of the Fe–C–TiO₂ photocatalysts in the process of phenol decomposition, some other photocatalysts were prepared with the residue carbon by mixing of FeC₂O₄ with TiO₂ and heating in Ar atmosphere at different temperatures, from 400 to 900 °C, and Fe–TiO₂ photocatalyst was prepared by heating FeC₂O₄ with TiO₂ at 550 °C in air. The characteristics of those samples are introduced in Table 3.

BET surface area was markedly reduced in those photocatalysts during the heating at high temperatures, what was caused by the sintering of TiO₂ particles. The content of carbon in Fe–C–TiO₂ photocatalysts was insignificant, ranged from 0.2 to 3.3 mass%. Transformation from anatase to rutile phase appeared at 700 °C. From the heat-treated temperature of 500 °C FeTiO₃ phase with ilmenite structure was clearly observed on XRD patterns of Fe–C–TiO₂ and Fe–TiO₂ photocatalysts. The highest content of iron measured by the extraction with 6 M HCl was noted for Fe–TiO₂ and FeC500–TiO₂ photocatalysts.

The adsorption of phenol on Fe–TiO₂ and Fe–C–TiO₂ with the residue carbon photocatalysts was performed before decomposition by the mixing of 0.1 g sample in 500 mL of phenol solution with concentration of 2.1×10^{-4} mol L⁻¹ in the dark for 3 h. These samples showed almost no adsorption of phenol.



Fig. 2. Phenol decomposition under UV and with H_2O_2 on Fe–TiO₂ and Fe–C–TiO₂ photocatalysts with residue carbon.

After adsorption, the process of phenol decomposition was performed in the same solution. In Fig. 2, the decomposition of phenol on Fe–TiO₂ and Fe–C–TiO₂ photocatalysts under UV irradiation with the addition of H₂O₂ is presented after saturation with adsorption. The initial phenol solution had pH 6.4, after adsorption pH did not change, after addition of H₂O₂ pH of solution dropped to 6.2 and with time of UV irradiation was insignificantly decreasing (after 1 h pH 5.3) due to the formation of decomposition products, and then was slightly increasing with following process of mineralisation (after 5 h pH 5.8).



Fig. 3. HPLC chromatograms recorded during phenol decomposition under UV and with addition of H_2O_2 on: (a) Fe–C500–TiO₂ and (b) Fe–TiO₂ photocatalysts.

Table 3 The characteristics of Fe–TiO₂ and Fe–C–TiO₂ with residual carbon

HTT (°C)	Sample code	BET (m ² /g)	Carbon (mass%)	Fe ³⁺ (mass%)	Fe ²⁺ (mass%)	Phases by XRD (A, anatase; R, rutile)
400	Fe-C400-TiO ₂	146	3.3	0.37	0.77	A
500	Fe-C500-TiO ₂	72	2.4	0.96	0.70	A, FeTiO ₃
600	Fe-C600-TiO ₂	28	0.9	0.07	0.33	A, FeTiO ₃
700	Fe-C700-TiO ₂	7.5	0.3	_	_	A, R, FeTiO ₃
800	Fe-C800-TiO ₂	3	0.3	0.06	0.09	R, FeTiO ₃
900	Fe-C900-TiO ₂	2.5	0.2	0.16	0.08	R, FeTiO ₃
550	Fe-TiO ₂	55	0	1.3	0.6	A, FeTiO ₃



Fig. 4. FT-IR spectra of: (a) Fe–PET400–TiO₂, (b) Fe–TiO₂ and (c) Fe–PET700–TiO₂ before and after 2 h of phenol decomposition under UV with addition of H_2O_2 .

The highest rate of phenol decomposition was observed on the Fe–C500–TiO₂ sample, which had the highest content of iron. Although, the content of iron in Fe–TiO₂ was also high, the phenol decomposition was going slowly, even much slower than on Fe–C–TiO₂ photocatalysts having only rutile phase (Fe–C800–TiO₂ and Fe–C900–TiO₂). It could be explained by the different mechanism of phenol decomposition. In Fig. 3, the chromatograms taken at different time of phenol decomposition under UV irradiation and with addition of H₂O₂ on Fe–TiO₂ and Fe–C500–TiO₂ photocatalysts are presented.

It can be observed that phenol decomposition on Fe–C500– TiO₂ proceeds progressively, phenol is very quickly converted to HQ, catehol, BQ and some unidentified short-chain products (peak observed at the retention time of 1.5 min), then ring products are decomposed completely after 2 h of UV irradiation and amount of short chain products decreases, they are completely decomposed after 5 h. On Fe–TiO₂ sample, however, phenol is also converted to HQ, catehol and short chain products (no BQ is observed), the decomposition is going slowly with no selectivity, all the organic compounds are decomposed gradually with time of UV irradiation, but also the highest concentration of HQ and catehol is observed after 2 h of UV irradiation.

Some FT-IR measurements of the chemical surface of $Fe-C-TiO_2$ and $Fe-TiO_2$ photocatalysts were performed before and after UV irradiation in the phenol solution with the addition of H_2O_2 . The results are presented in Fig. 4.

The chemical surface of Fe–PET400–TiO₂ photocatalyst changed after 2 h of phenol decomposition under UV with H_2O_2 , some O–H and C=O bands reasonably attributed to hydroquinones and benzoquinone were formed (O–H band became wider), whereas on Fe–TiO₂ only C=O band appeared, which could be attributed to benzoquinone, and no change of the chemical surface of Fe–PET700–TiO₂ photocatalyst was observed.

4. Discussion

In the photo-Fenton process, the degradation of organic compounds proceeds by the formation of OH[•] radicals due to the oxidation and reduction of iron. In the presence of Fe^{2+} and H_2O_2 , the following reaction occurs:

$$Fe^{2+} + H_2O_2 \rightarrow Fe^{3+} + OH^{\bullet} + OH^{-}$$
(1)

Iron can be reduced in the presence of UV light and hydroxyl ions, according to the reaction:

$$Fe^{3+} + OH^- + h\nu \rightarrow Fe^{2+} + OH^{\bullet}$$
⁽²⁾

Reaction (2) is less efficient than reaction (1), so when the iron exists only in the form of Fe^{3+} , then the process of decomposition is slowed down until buildup of Fe^{2+} [12].

From FT-IR measurements it could be concluded that formed hydroquinones, which were adsorbed on the surface of Fe–C–TiO₂ photocatalyst during phenol decomposition played important role in the photo-Fenton reaction and could result in the acceleration of the phenol decomposition. Therefore, the contact of phenol molecules with the photocatalyst surface is important, in case of Fe–C–TiO₂ photocatalysts with high carbon content which showed poor adsorption of phenol, the process of phenol decomposition was accelerated in the beginning, because of the oxidation of Fe²⁺ in the presence of H₂O₂, and then was slowed down due to the slow reduction of Fe³⁺, what could be explained by the poor contact of formed hydroquinones with iron (Fig. 1).

On Fe–C–TiO₂ photocatalysts with the residue carbon, phenol decomposition under UV and with H_2O_2 was going faster in case of the photocatatalysts, which had the higher content of iron, especially Fe²⁺ (Fig. 2), the speed of phenol decomposition was determined mostly by the occurring photo-Fenton reactions.

On Fe–TiO₂ photocatalyst decomposition of phenol under UV with H_2O_2 was accelerated in comparison with pristine TiO₂, however was going much slower than on Fe–C–TiO₂ photocatalysts, as shown in Fig. 2. In that case phenol was decomposed by the radical reaction, which was not selective for phenol. Therefore, the presence of carbon in Fe–C–TiO₂ photocatalysts seems to be important, because on Fe–C–TiO₂ photocatalysts the phenol conversion was going quickly to ring products, which took part in the further mineralization of phenol.

5. Conclusions

High rate of phenol decomposition was obtained on Fe–C–TiO₂ photocatalyst under UV irradiation with addition of small amount of H_2O_2 to the phenol solution, because some photo-Fenton reactions worked for phenol decomposition. Decomposition of phenol on Fe–C–TiO₂ photocatalysts proceeded mainly by the photo-Fenton process, oxidation of Fe²⁺ to Fe³⁺ in the presence of H_2O_2 and the reduction of Fe³⁺ under UV and in the presence of some hydroquinones. It was proved that the rate of phenol decomposition was governed by

two factors, high content of iron in the photocatalyst (with a good balance between Fe^{2+} and Fe^{3+}), and a good contact of phenol molecules with the active iron species. In case of the high content of carbon (7.5–15.5 mass%) in Fe-C-TiO₂ photocatalysts (comes from the carbonization of PET and oxalate), adsorption of phenol onto the surface of Fe-C-TiO2 is important to facilitate contact of phenol molecules with the active species formed during excitation of semiconductor and photo-Fenton reactions. In case of the low content of carbon in Fe–C–TiO₂ photocatalysts (0.2-3.3 mass%), the rate of phenol decomposition is higher for the higher content of the active iron in the photocatalyst. On Fe-TiO₂ photocatalyst, decomposition of phenol under UV irradiation and with addition of H₂O₂ is going faster than on the pristine TiO₂ due to the photo-Fenton reactions, but much slower than on the Fe–C–TiO₂ photocatalysts modified by iron, which contain the carbon. On TiO₂ and Fe-TiO₂ photocatalysts phenol is decomposed by the radical reaction, which is not selective for phenol decomposition. From those measurements, it was proved that phenol decomposition is much favored by the complex reactions of phenol oxidation and it could be realized by photo-Fenton reactions with using Fe-C-TiO₂ photocatalysts.

Acknowledgement

The present work was supported by the Research Project of Environmental Technology Development Fund.

References

- T. Tsumura, N. Kojitani, I. Izumi, N. Iwashita, M. Toyoda, M. Inagaki, J. Mater. Chem. 12 (2002) 1391.
- [2] T. Tsumura, N. Kojitani, H. Umemura, M. Toyoda, M. Inagaki, Appl. Surf. Sci. 196 (2002) 429.
- [3] B. Tryba, A.W. Morawski, T. Tsumura, M. Toyoda, M. Inagaki, J. Photochem. Photobiol. A: Chem. 167 (2004) 127.
- [4] M. Inagaki, H. Miura, H. Konno, J. Eur. Ceram. Soc. 18 (1998) 1011.
- [5] M. Toyoda, Y. Yoshikawa, T. Tsumura, M. Inagaki, J. Photochem. Photobiol. A: Chem. 7 (2005) 67.
- [6] M. Inagaki, F. Kojin, B. Tryba, Carbon 43 (2005) 652.
- [7] B. Tryba, M. Toyoda, A.W. Morawski, M. Inagaki, Chemosphere 60 (2005) 477.
- [8] M. Pera-Titus, V. Garcia-Molina, M.A. Baños, J. Giménez, S. Esplugas, Appl. Catal. B: Environ. 47 (2004) 219.
- [9] M. Neamtu, A. Yediler, I. Siminiceanu, A. Kettrup, J. Photochem. Photobiol. A: Chem. 161 (2003) 87.
- [10] M.I. Franch, J.A. Ayllón, J. Peral, X. Doménech, Appl. Catal. B: Environ. 50 (2004) 89.
- [11] H. Fallmann, T. Krutzler, R. Bauer, S. Malato, J. Blanco, Catal. Today 54 (1999) 309.
- [12] R. Chen, J.J. Pignatello, Environ. Sci. Technol. 31 (1997) 2399.
- [13] F. He, L. Lei, J. Zhejiang Univ. Sci. 5 (2004) 198.
- [14] R. Bauer, G. Waldner, H. Fallmann, S. Hager, M. Klare, T. Krutzler, S. Malato, P. Maletzky, Catal. Today 53 (1999) 131.
- [15] H. Hidaka, H. Honjou, S. Horkoshi, N. Serpone, New J. Chem. 27 (2003) 1371.
- [16] N. Majcen, Acta Chim. Slov. 45 (1998) 85.